



# International Journal of Fisheries and Aquatic Studies

E-ISSN: 2347-5129

P-ISSN: 2394-0506

(ICV-Poland) Impact Value: 5.62

(GIF) Impact Factor: 0.549

IJFAS 2018; 6(2): 621-627

© 2018 IJFAS

www.fisheriesjournal.com

Received: 11-01-2018

Accepted: 12-02-2018

**Popoola Omoniyi Michael**

Department of Fisheries and  
Aquaculture Technology,  
Federal University of  
Technology, Akure Nigeria

## Sublethal effect of diuron on DNA status, hematological and biochemical indices of fresh water catfish, *Clarias gariiepinus* Juveniles

**Popoola Omoniyi Michael**

### Abstract

The toxicity of Diuron (N-(3, 4 dichlorophenyl)-N, N- dimethylurea), a herbicide was investigated on *Clarias gariiepinus* popularly cultured fish in Nigeria. Hematological and biochemical indices status of DNA were examined in Diuron treated fish. Juvenile African catfish (*Clarias gariiepinus*) (76.69±2.85g) were exposed to sub-lethal 0.00 ml/L, 0.5 ml/L, 0.75 ml/L, 1.00 ml/L, 1.25 ml/L and 1.50 ml/L of herbicide for 96 hours under static bioassay conditions. Another group of 2mg/l of cyclophosphamide was prepared. Water quality parameters were monitored during the bio-assay test. After 96 hours, blood was extracted from the fish and haematological parameters (Hb, PCV, WBC and RBC) were estimated. Biochemical indices of changes in creatinine, urea, cholesterol and triglyceride level were determined. The genotoxic effect was evaluated using micronucleus test. The median lethal concentration (LC<sub>50</sub>) value was determined to be 1.29ml/L. There were significant differences (p<0.05) in the haematological, biochemical analysis as the parameters measured differed with concentrations of the toxicant. The induction of micronuclei was highest (15.67) with highest concentration of the toxicant in the peripheral blood. The study revealed that long exposure to diuron at sublethal concentration can cause mutagenic effects, alter blood chemistry and oxidative dysfunctions in *C. gariiepinus*.

**Keywords:** diuron; sub-lethal toxicity; genotoxicity; haematological indices; biochemical indices; *Clarias gariiepinus*; sub-lethal concentration

### Introduction

Pollution in aquatic environment affect both flora and fauna which eventually result in disease, destruction of vital organs, reduces growth as well as reproduction performance and death of the aquatic organism when it is beyond the tolerant level of the organism (Beneet *et al.*, 1995 and Rahaman *et al.*, 2002)<sup>[1, 2]</sup>.

Herbicides have adverse effects on the environment and persist in the environment long after they have carried out their weed killing function. Their persistence in the soil is a public health concern because of the potential of leaching into ground water and contaminating surface water.

Fishes and other aquatic organisms are directly affected by pollution caused by herbicides consequently the effect on humans is indirect because they are the final consumer of these aquatic products. Usually, there is a wide spread exposure of humans to herbicides, especially during rainy season as a result of their use for land preparation by famers, and elimination of weeds and unwanted vegetation by households. Hence, the use of mortality or bioassay experiments is necessary to assess the environmental impact of toxic compounds.

Diuron (N-[3, 4-dichlorophenyl]-N, N-dimethylurea), A herbicide of the phenylurea family, is widely used to destroy unwanted weeds and as an antifouling agent in paints. Diuron is generally persistent in soil, water and groundwater. It is also slightly toxic to mammals and birds as well as moderately toxic to aquatic invertebrates depending on the concentration (Giacomazzi and Cochet, 2004)<sup>[3]</sup>.

The fish haematology status is affected by the use of various herbicides whether selective or non-selective in which the effect may be acute or chronic. Changes in blood properties in response to environmental conditions are a response to environmental stress and can be considered as an important bio-indicator. Evaluation of the haematological profile usually gives information on the response of the fish to injury, size, weight and stress (Nussey *et al.*,

### Correspondence

**Popoola Omoniyi Michael**

Department of Fisheries and  
Aquaculture Technology,  
Federal University of  
Technology, Akure Nigeria

1995)<sup>[4]</sup>. Blood parameters are therefore considered as pathophysiological indicators of the whole body and are used in diagnosing the structural and functional status of fish exposed to various toxicants (Adhikari *et al.*, 2004; Maheswaran *et al.*, 2008)<sup>[5, 6]</sup>.

The cytogenetic analysis, such as micronucleus induction, chromosomal aberrations and sister chromatid exchange offers reliable genetic assays in detecting genotoxic chemicals at sub-toxic levels. The micronucleus is a cytoplasmic chromatin mass with a small nucleus produced from a lagging chromosome in the anaphase stage as a result of some structural and/or numerical chromosomal aberrations in the cells during mitosis (Parveen and Shadab, 2012)<sup>[7]</sup>. This aberration is due to a genetic imbalance in the cells, possibly leading to carcinogenesis (Tolga Cavas *et al.*, 2005)<sup>[8]</sup>.

There several biological test used in detection and identification of environmental pollutants (Güner and Muranlı, 2011)<sup>[9]</sup>. The use of micronuclei (MN) count has been reported by Bombail (2011)<sup>[10]</sup> an indices of chromosome aberration and non- functioning of spindle in mitosis. The benefits of this test involve its simplicity, reliability, and sensitivity.

Fish are considered good candidate for monitoring aquatic genotoxicity due to their ability to metabolise xenobiotics and accumulate pollutants (Grisolia and Corderio, 2000)<sup>[11]</sup>. Effects of Malathion on *Chana punctatus* has been reported by Parveen and Shadab (2011)<sup>[12]</sup> using the micronucleus test Mohamed *et al.* (2008)<sup>[13]</sup> analysed the cytogenetic damage by measuring the chromosomal aberration in the gill cells.

It is therefore necessary to know the effect of diuron herbicide on the juvenile of *Clarias gariepinus* bearing in mind that these juveniles grow to big size for human consumption.

This research therefore designed to examine the mutagenic and haematological effects as well as antioxidant stress biomarkers *C. gariepinus* after long term exposure to diuron herbicide.

## Materials and Methods

### Experimental Fish and Toxicant

A total 250 of *C. gariepinus* (mean weight, 76.69±2.85g) were collected between July and August 2017, acclimated for 72 hours in the Department of Fisheries and Aquaculture Research Farm of the Federal University of Technology, Akure. Diuron was obtained from Agro Shop in Akure Ondo State. The stock solution of Diuron was prepared by dissolving 150ml in 1 litre of distilled water from the different static toxicity test doses were calculated and prepared by appropriate dilution.

### Experimental Procedure

Five concentrations (0.5, 0.75, 1.0, 1.25, 1.5 ml) of Diuron were prepared and tested on the *Clarias gariepinus* juveniles for the static bio-assay test together with another group containing 2mg/l of cyclophosphamide. A total of 14 glass tanks (0.7 x 0.5 x 0.45m<sup>3</sup>) were used (5 treatment and 2 control groups) containing 10 litres of water in each was used for the dilution of each concentration. The exposure was continued for 96 hour during which physicochemical parameters of the test media were recorded. The physicochemical parameter measured and recorded during the experiment were temperature, total dissolved solid, conductivity and salinity using EXTECH Instrument EC 500 and the dissolved oxygen was measured using a Digital dissolved oxygen metre model 831E.

Mortality rate was also observed at 24, 48, 72 and 96 hours respectively. Confirmation of dead fish was done by gentle touch with a glass rod and by noticing no opercula movement and tail-fin beats. Dead fishes were promptly removed and mortality was recorded. The median lethal concentration (LC<sub>50</sub>) at 96 h was computed using the probit analysis.

The blood samples from the life fish were collected after 96 hours for further study.

### Haematological examination

Blood samples were collected from fishes by puncturing caudal vein utilizing ethylenediaminetetraacetate (EDTA) as anticoagulant. The blood, (2.5 ml) was emptied into heparinized bottles for determination of hematocrit. Hemoglobin (Hb) fixation was estimated with Hb test unit utilizing the cyanmethemoglobin technique. Red blood cell count (RBC) and white blood cell count (WBC) count were checked under light microscope with an enhanced Neubauer haemocytometer using method of Shah and Altindag, (2005)<sup>[14]</sup>. The mean corpuscular volume (MCV), mean corpuscular hemoglobin (MCH) and mean corpuscular hemoglobin fixation (MCHC) were computed in picograms = Hb/RBC x 10; and MCHC = (Hb in 100 mg blood/Hct) x 100.

### Biochemical examination

The urea nitrogen was estimated according to Randox laboratories limited procedure (Weatherburn, 1967)<sup>[15]</sup>. The creatinine colorimetric assay was conducted according to Perone *et al.*, (1992)<sup>[16]</sup>. The total cholesterol concentration was determined using colorimetric method by Allain *et al.*, (1974)<sup>[17]</sup> with the standard formula

$$\text{Cholesterol conc.} = \frac{A_{\text{serum}}}{A_{\text{standard}}} \times n$$

The triglyceride concentration was determined colorimetrically using method of Tietz (1982)<sup>[18]</sup> with

$$\text{Triglyceride conc.} = \frac{A_{\text{serum}}}{A_{\text{standard}}} \times n$$

### Micronucleus assay

The method described by Ayllon and Garcia-Vazquez (2000)<sup>[19]</sup> for the piscine micronucleus test was used with little modification. Blood samples from five (5) selected fish from each of the sub-lethal concentration of diuron were smeared on clean, grease free iced glass slides. The treated slides were embedded in methanol for 12 min and later allow to air-dry at room temperature. The slides were again stained with 6% Geimsa in Sorenson buffer (pH 6.9) for 25 min. Graded alcohol was used for the dehydration of the slides followed by clearing in xylene solution. A mixture of distyrene, plasticizer and xylene were used in mounting the slides. From each of the slides, about 1000 erythrocyte cells were scored under a light microscope (Leica DM300). In identifying erythrocyte cell, each MN were subjected to the same color, plane of focus, clearly separated and smaller than one-third of the main nucleus as described by Nwani *et al.* (2011)<sup>[20]</sup>.

### The MN frequency was calculated thus:

$$\text{MN\%} = \frac{\text{Number of cells containing micronuclei}}{\text{Number of cells containing micronuclei} \times 100}$$

### Data analysis

Values from the Haematological parameters, water quality

parameter and Biochemical indices were analysed by one factor analysis using test Statistical Package for Social Science (SPSS 23) and the difference in the mean was separated using the New Duncan Multiple Range Test (NDMRT). The mortality rate was determined using probit analysis. The frequency of micronuclei was presented using chart.

**Results**

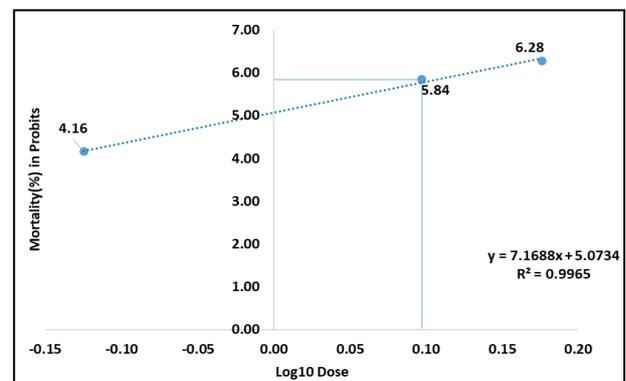
The probit analysis of the log concentration against the probit percentage mortality of the fish exposed to diuron herbicide for 96 hours was presented in figure 1. In the control tank, the experimental fishes exhibited normal behavioural and swimming patterns compared with other treatments. There was a reduction in dissolved oxygen concentration, increases in conductivity and total dissolved solid level with increase in concentration and time Table 1.

**Haematological Parameters**

The haematological properties (PVC, HB, WBC, RBC, Neutrophyl, Monocytes Lymphocyte MCV MCHC and MCH) of the exposed fish at different concentrations were obtained after 96 hours exposure and shown in the table 2  
 There was a significant difference ( $p < 0.05$ ) in Packed cell volume (Haemtocrit) and was noticed to range from  $10.5 \pm 0.95$  to  $16.80 \pm 0.95$ , which reduced with concentration of the toxicant. Haemoglobin exhibit significant different ( $p < 0.05$ ), its values ranged from  $4.40 \pm 0.22$  to  $5.85 \pm 0.22$  and decreased with increase in concentration of the Diuron.  
 There was a significant difference ( $p < 0.05$ ) in WBC with values ranged from  $2075 \pm 434.58$  to  $4850 \pm 434.58$ , and WBC showed a decrease in Treatment 1 (0.5ml/l) to Treatment 5 (1.50ml/l) with increase in concentration compared with the control. Red blood cell (RBC) ranged from  $1.3 \pm 0.43$  to  $3.95 \pm 0.43$ , and decreased with increase in concentration when compared with the control. Neutrophyl values ranged from  $4.60 \pm 0.61$  to  $8.55 \pm 0.61$ , the neutrophyl increased in

Treatment 1, 2, 3, 4, 5 (0.5ml/l, 0.75ml/l, 1.0ml/l, 1.25ml/l, 1.50ml/l) with increase in concentration compared with the control. Monocytes ranged from  $23.60 \pm 0.54$  to  $26.92 \pm 0.54$ , which showed an increase in treatments with increase in concentration when compared with the control. Lymphocytes ranged from  $68.10 \pm 0.28$  to  $69.86 \pm 0.28$ , and decreased in Treatments 1, 2, 3, 4, 5 (0.5ml/l, 0.75ml/l, 1.0ml/l, 1.25ml/l, 1.50ml/l) respectively when compared with the control with increase in concentration. MCV ranges from  $37.15 \pm 1.77$  to  $88.06 \pm 21.95$ , and decreased in Treatments 1 and 2 (0.5ml/l, 0.75ml/l), and increase Treatment 3, 4, 5 (1.0ml/l, 1.25ml/l, 1.50ml/l) with increases in concentration when compared with the control.

MCH ranged from  $14.33 \pm 1.18$  to  $37.40 \pm 10.71$ , the MCH value obtained showed a decrease in Treatment 1, 2 (0.5ml, 0.75ml), and an increase in Treatment 3, 4, 5 (1.0ml, 1.25ml, 1.50ml) respectively compared with the control with increase in concentration. MCHC result ranged from  $34.86 \pm 0.64$  to  $43.88 \pm 1.80$ , the MCHC of the treatments increase with increase in concentration of Diuron compared with the control.



**Fig 1:** Lethal concentration (LC50) of *Clarias gariepinus* exposed to Diuron after 96 hou

**Table 1:** Physicochemical Parameter of water recorded 96 hour Bio-assay test

| Parameter            | 0.00                        | 0.50                        | 0.75                        | 1.0                         | 1.25                        | 1.50                        |
|----------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|-----------------------------|
| Conductivity (µS/cm) | 231.00 ± 12.00 <sup>a</sup> | 250.00 ± 22.00 <sup>a</sup> | 252.50 ± 15.50 <sup>a</sup> | 282.00 ± 14.00 <sup>a</sup> | 428.50 ± 15.50 <sup>b</sup> | 403.50 ± 37.50 <sup>b</sup> |
| TDS (mg/L)           | 112.35 ± 5.95 <sup>a</sup>  | 121.80 ± 14.85 <sup>a</sup> | 123.20 ± 7.70 <sup>a</sup>  | 137.90 ± 7.00 <sup>a</sup>  | 210.00 ± 7.00 <sup>b</sup>  | 197.40 ± 18.20 <sup>b</sup> |
| Salinity (g/kg)      | 0.05 ± 0.00 <sup>a</sup>    | 0.06 ± 0.00 <sup>a</sup>    | 0.06 ± 0.00 <sup>a</sup>    | 0.07 ± 0.00 <sup>a</sup>    | 0.10 ± 0.00 <sup>b</sup>    | 0.10 ± 0.00 <sup>b</sup>    |
| pH (mol / L)         | 7.81 ± 0.09 <sup>b</sup>    | 7.63 ± 0.03 <sup>a</sup>    | 7.72 ± 0.03 <sup>ab</sup>   | 7.74 ± 0.00 <sup>ab</sup>   | 7.84 ± 0.04 <sup>b</sup>    | 7.85 ± 0.01 <sup>b</sup>    |
| Temperature (°C)     | 23.95 ± 0.05 <sup>a</sup>   | 23.95 ± 0.05 <sup>a</sup>   | 23.90 ± 0.00 <sup>a</sup>   | 23.95 ± 0.05 <sup>a</sup>   | 23.95 ± 0.05 <sup>a</sup>   | 23.95 ± 0.05 <sup>a</sup>   |
| DO (mg/l)            | 5.65 ± 0.05 <sup>e</sup>    | 3.30 ± 0.10 <sup>d</sup>    | 3.05 ± 0.05 <sup>c</sup>    | 2.65 ± 0.05 <sup>b</sup>    | 2.15 ± 0.05 <sup>a</sup>    | 2.00 ± 0.00 <sup>a</sup>    |

\*TDS= Total dissolved solid \*DO= Dissolved oxygen

The values are expressed as the mean ± standard error. Means in the same horizontal column followed by different superscript are significantly different ( $p < 0.05$ ).

**Table 2:** Haematological parameters of *C. gariepinus* exposed Diuron Concentrations

|                                          | Concentrations (ml/L)      |                             |                              |                             |                            |                            |
|------------------------------------------|----------------------------|-----------------------------|------------------------------|-----------------------------|----------------------------|----------------------------|
|                                          | 0.00                       | 0.50                        | 0.75                         | 1.00                        | 1.25                       | 1.50                       |
| PCV (%)                                  | 16.80 ± 0.95 <sup>b</sup>  | 13.67 ± 0.95 <sup>ab</sup>  | 12.75 ± 0.95 <sup>a</sup>    | 12.00 ± 0.95 <sup>a</sup>   | 10.75 ± 0.95 <sup>a</sup>  | 10.5 ± 0.95 <sup>a</sup>   |
| HB(g/100ml)                              | 5.85 ± 0.22 <sup>c</sup>   | 5.40 ± 0.22 <sup>bc</sup>   | 4.75 ± 0.22 <sup>ab</sup>    | 4.70 ± 0.22 <sup>ab</sup>   | 4.70 ± 0.22 <sup>ab</sup>  | 4.40 ± 0.22 <sup>a</sup>   |
| WBC (×10 <sup>3</sup> /mm <sup>3</sup> ) | 4850 ± 434.58 <sup>c</sup> | 4600 ± 434.58 <sup>bc</sup> | 3600 ± 434.58 <sup>abc</sup> | 3050 ± 434.58 <sup>ab</sup> | 2875 ± 434.58 <sup>a</sup> | 2075 ± 434.5 <sup>a</sup>  |
| RBC (×10 <sup>6</sup> /mm <sup>3</sup> ) | 3.95 ± 0.43 <sup>c</sup>   | 3.7 ± 0.43 <sup>c</sup>     | 3.35 ± 0.43 <sup>bc</sup>    | 2.75 ± 0.43 <sup>abc</sup>  | 1.85 ± 0.43 <sup>ab</sup>  | 1.3 ± 0.43 <sup>a</sup>    |
| Neutrophyl (%)                           | 4.60 ± 0.61 <sup>a</sup>   | 5.2 ± 0.61 <sup>ab</sup>    | 5.12 ± 0.61 <sup>ab</sup>    | 6.47 ± 0.61 <sup>abc</sup>  | 7.21 ± 0.61 <sup>bc</sup>  | 8.55 ± 0.61 <sup>c</sup>   |
| Monocytes (%)                            | 23.60 ± 0.54 <sup>a</sup>  | 24.20 ± 0.54 <sup>a</sup>   | 25.50 ± 0.54 <sup>ab</sup>   | 26.92 ± 0.54 <sup>b</sup>   | 24.10 ± 0.54 <sup>a</sup>  | 26.18 ± 0.54 <sup>b</sup>  |
| Lymphocyte (%)                           | 69.86 ± 0.28 <sup>c</sup>  | 68.34 ± 0.28 <sup>ab</sup>  | 69.20 ± 0.28 <sup>bc</sup>   | 68.33 ± 0.28 <sup>ab</sup>  | 68.10 ± 0.28 <sup>a</sup>  | 69.20 ± 0.28 <sup>bc</sup> |
| MCV(fl)                                  | 42.78 ± 2.27 <sup>a</sup>  | 37.15 ± 1.77 <sup>a</sup>   | 38.33 ± 2.11 <sup>a</sup>    | 44.18 ± 3.49 <sup>a</sup>   | 60.19 ± 8.93 <sup>ab</sup> | 88.06 ± 21.95 <sup>a</sup> |
| MCHC(gm/100)                             | 34.86 ± 0.64 <sup>a</sup>  | 39.58 ± 1.12 <sup>abc</sup> | 37.33 ± 1.03 <sup>ab</sup>   | 39.27 ± 1.25 <sup>abc</sup> | 43.88 ± 1.80 <sup>c</sup>  | 42.06 ± 1.68 <sup>bc</sup> |
| MCH(gm/100)                              | 14.93 ± 1.07 <sup>a</sup>  | 14.73 ± 1.12 <sup>a</sup>   | 14.33 ± 1.18 <sup>a</sup>    | 17.39 ± 1.92 <sup>a</sup>   | 26.57 ± 5.00 <sup>ab</sup> | 37.40 ± 10.71 <sup>b</sup> |

Values are mean ± standard error. Means in the same column followed by different superscripts are significantly different ( $p < 0.05$ ). PCV-Packed cell volume, HB- Haemoglobin, WBC-White blood cell, RBC-Red blood cell, MCV-Mean cell volume, MCH- Mean cell haemoglobin, MCHC-Mean cell haemoglobin concentration

### Biochemical Examination

The result of biochemical parameters observed in this study is presented in table 3. There was a significant difference ( $p < 0.05$ ) in all the parameters measured. The result showed that the creatinine, urea, triglyceride and total cholesterol level increased with increase in the concentration of diuron.

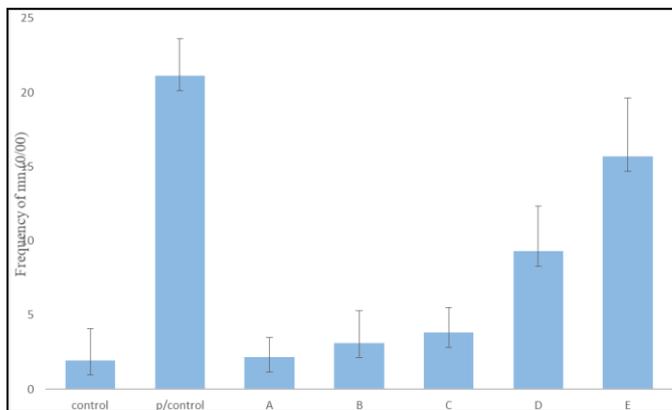
### Induction of Micronucleous

Figure 2 shows induced MN frequencies in red blood cell C.

**Table 3:** Biochemical Composition of *C. gariepinus* exposed to different concentration of Diuron Concentration ml/l

| Parameter         | 0                       | 0.5                      | 0.75                      | 1.00                      | 1.25                     | 1.50                     |
|-------------------|-------------------------|--------------------------|---------------------------|---------------------------|--------------------------|--------------------------|
| Creatinine        | 4.22±0.55 <sup>a</sup>  | 5.01±0.12 <sup>a</sup>   | 6.05±0.18 <sup>b</sup>    | 6.66±0.18 <sup>b</sup>    | 7.70±0.12 <sup>c</sup>   | 8.19±0.25 <sup>c</sup>   |
| Urea              | 52.50±2.50 <sup>a</sup> | 97.50±7.50 <sup>b</sup>  | 137.50±2.50 <sup>c</sup>  | 157.50±2.50 <sup>d</sup>  | 162.50±2.50 <sup>d</sup> | 197.50±2.50 <sup>e</sup> |
| Triglyceride      | 54.36±2.18 <sup>a</sup> | 81.54±3.27 <sup>b</sup>  | 100.02±4.35 <sup>c</sup>  | 121.77±4.35 <sup>d</sup>  | 139.15±4.35 <sup>e</sup> | 161.99±5.44 <sup>f</sup> |
| Total Cholesterol | 98.70±0.42 <sup>a</sup> | 101.22±1.26 <sup>a</sup> | 107.10±2.94 <sup>ab</sup> | 113.40±2.52 <sup>bc</sup> | 120.12±3.36 <sup>c</sup> | 133.56±2.52 <sup>d</sup> |
| Lactate           | 0.56±0.47 <sup>a</sup>  | 1.45±0.47 <sup>ab</sup>  | 2.19±0.47 <sup>abc</sup>  | 2.77±0.47 <sup>bc</sup>   | 3.22±0.47 <sup>c</sup>   | 3.59±0.47 <sup>c</sup>   |

Values were derived in duplicate with significant difference observed across all the treatment ( $p < 0.05$ ). The values are expressed as the mean  $\pm$  standard error. Means in the same horizontal column followed by different superscript are significantly different



**Fig 2:** Frequencies (%o) of MNs, red blood cells from *C. gariepinus* exposed to the Diuron herbicide for 96 hours Cyclophosphamide (2 mg/L) was as positive control 0.0 ml/l of diuron as negative control

### Discussion

Diuron, an herbicide used for specific crops and as an algacide in ornamental aquaria is toxic to fish and poses a threat to the health status of *Clarias gariepinus* juveniles. Increased concentration of Diuron could impair the biochemical indices of the fish. In terms of behavioral patterns, various behavioural response such as gulping air at the surface, abnormal movement was observed and the observations agreed with that of Lin and Liu, (1990) [21] who reported that clinical signs such as abnormal movement and high respiration rate in hybrid tilapia (*Oreochromis mossambicus*) induced by ammonia suggested neurological dysfunction and gill damage. Likewise, the deficiency of oxygen causes the hypoxic condition in fish which results in an increase in the breathing rate and to cope with the condition, the fishes gulp air by frequent surfacing. The test fish exhibited somersaulting activity, erratic swimming, swimming upside down and with time settled at the bottom of the tank and almost 95% died before the end of the exposure period. In the other concentrations the fish were seen to come to the water surface to gulp air, high rate fin and opercula movement. This is in agreement with the work of Omoniyi *et al.* (2002) [22]; Kulakkattolick and Kramer (1997) [23] who reported several abnormal behaviours such as incessant jumping and gulping of air, restlessness, surface to bottom movement, sudden quick movement and resting at the bottom also observed that the fish became inactive at higher

*gariepinus* exposed to Diuron for 96 hours. The highest value was recorded in concentration 1.50ml/l (15.67%) and least in 0.5ml/l (2.17ml/l) for 96 hours period. There was a reduction in MN induction frequency in the blood erythrocytes of the *Clarias gariepinus* and the induction of MN is observed to be concentration – dependent as the percentage of induction increased with exposure time.

concentrations with increasing time of exposure to toxicant which is a normal observation in acute and chronic toxicity test.

The physicochemical properties measured in the test solutions showed that water temperature and pH varied slightly between the treatments; there was also decrease in the dissolved oxygen level and increase in conductivity and total dissolved solids and this agrees with the findings of Yunis *et al.*, (2014) [24] who reported that fish mortality could not be attributed to pH; but instead were associated with the low levels of DO when the neotropical fish *Hyphessobrycon eques* was exposed to the aqueous extract of *Uncaria tomentosa* bark. The increase in the toxicant may be the cause of the depletion of oxygen which agrees with Ukwagu *et al.*, (2012) [25] who stated that the physicochemical parameters examined from the tank polluted with effluent showed a decline in Dissolved oxygen level and increase in total dissolved solid and conductivity level

According to Barkhordar (2013) [26] the 96 hrs LC<sub>50</sub> tests are conducted to measure the susceptibility and survival potential of organisms to particular toxic substances. The result obtained from this investigation revealed that the 96-hour LC<sub>50</sub> for the juvenile of *Clarias gariepinus* exposed to Diuron was 1.25ml/L.

The fish in the highest concentration 1.50ml/l showed 95% mortality while the fish in the lower concentrations of (0.5ml/l and 0.75ml/l) showed 0% mortality which showed that stress and eventual death of the fish is concentration dependent. The mortality rate was higher in the higher concentrations and this is in agreement with Fryer (1997) [27] who observed that the highest concentration of the toxicant resulted in the highest mortality rate which was described as normal, that in all toxicant thresholds is reached above which there is no drastic survival of animal. Below the threshold, animal is in a tolerance zone, above the tolerance zone is the zone of resistance.

In this study the result on the haematological parameters of *Clarias gariepinus* juveniles exposed to different concentrations of the toxicant showed a significant difference in the blood parameter when compared to the control. The result showed that there was a decrease in the PCV level of the test fish with an increase in the concentration; the fish subjected to the highest concentration (1.50ml/l) had the lowest PCV level although the decrease in packed cell volume

indicated that the fish might be suffering from anaemia. This corroborated the finding of (Oropesa *et al.*, 2009) <sup>[28]</sup> where fishes exposed to atrazine experienced decreases in the packed cell volume levels. The haemoglobin level decreased with increase in concentration, the fish subjected to the highest concentration (1.5ml/l) had the lowest haemoglobin level. The reduction in haemoglobin concentration in experimental fish might be due to the destruction of red blood cell as reported by (Svoboda *et al.*, 2001) <sup>[29]</sup> where he stated that exposure to heavy metals or pesticides leads to reduced haemoglobin content and haematocrit via disorders in haemopoietic and accelerated disintegration of erythrocyte cell membrane. The decrease in blood haemoglobin levels as seen in this research suggests a response of the juvenile fish to decrease in the blood oxygen carrying capacity. As haemoglobin level decreases, PCV level decreases as well. This is because haemoglobin will not decrease without a corresponding decrease in the PCV value which indicate a negative absorption of the toxicant in contrast with Olufayo (2009) <sup>[30]</sup>; Ayotunde *et al.*, (2004) <sup>[31]</sup> who explained that exposure of *C. gariepinus* to sub-lethal concentrations of *Derris elliptica* caused a significant increase in PCV, haemoglobin, and erythrocyte of the fish treated with high concentration of *Derris elliptica*.

The decrease in blood haemoglobin and red blood cells may also be due to the presence of stressor (diuron) which causes haemodilution to occur as a result of impaired osmoregulation (Rottman *et al.*, 1992) <sup>[32]</sup>. There was reduction in the red blood cells with increase in concentration in comparison with the control group which suggested that at higher concentrations, there could have been destruction of red blood cells. Similar patterns were also observed for haemoglobin (Hb) values which suggested haemodilution which could lead to anaemia. The swelling of erythrocytes can be inferred from the increase in the MCV, which is an indicator of the size or state of the red blood cells which agrees with Larsson *et al.*, (1985) <sup>[33]</sup>. who attributed the increase in MCV to swelling of the red blood cells due to hypoxic conditions or impaired water balance (Osmotic stress) and this may be related to macrocytic anaemia in fishes exposed to metal pollution. Ayotunde *et al.*, (2004) <sup>[31]</sup>; Adekunle *et al.*, (2007) <sup>[34]</sup> stated that the haematological examination of the fingerlings showed an increase in RBC, WBC, PCV, Hb, MCH and MCHC values, while there was a reduction in the ESR, MCH and MCV values and lymphocytes percentage showed significant increase. In this study the white blood cell which serve as the immune response decreased significantly from the control group there was significant difference between treatments showing that the toxicant was beyond the immune level of the fish such that it caused destruction of the white blood cell. Mean corpuscular haemoglobin (MCH) measures the average amount of haemoglobin within a red cell show increase in values with concentrations increase. In accordance to this finding, (Ruperelia, 1990) <sup>[35]</sup> reported that MCHC levels increased in tilapia *Oreochromis mosambicus* when it was exposed to cadmium.

Creatinine is a nitrogenous waste product that is eliminated by the kidney when excretion is suppressed in renal inadequacy. In this research, the creatinine value of the fish exposed to diuron increased significantly with increase in toxicant concentration. Murray *et al.*, (1990) <sup>[36]</sup> and Lall *et al.*, (1997) <sup>[37]</sup> reported that elevated creatinine in serum might be induced by glomerular inefficiency and a rise in creatinine value is an indication of renal tubular damage. Thus, it could

be inferred from this work that diuron is nephrotoxic. Triglyceride measured in this study showed a significant increase with increase in concentration. The increase in blood triglyceride and cholesterol may be due to dysfunction of liver and destruction of cell membranes can also lead to increased levels of cholesterol in plasma. Disorder in triglyceride uptake by adipose tissue may also increase triglycerides. John (2007) <sup>[38]</sup> explains cholesterol and glucose levels in blood of freshwater fish, *Mystus vittatus*, increased after exposure to metasytox and sevin. Increase of stress hormones such as cortisol in blood of fish exposed to various insecticides, which stimulates lipid breakdown in adipose tissue, were also found in Banae (2013) <sup>[39]</sup>. Similar changes have also been reported by Ibrahim and El-Gamal (2003) <sup>[40]</sup>, Lasram *et al.* (2009) <sup>[41]</sup> and Acker and Nogueira (2012) <sup>[42]</sup>.

Genotoxic assays such as MN test had been reported by Mazzeo and Marin-Morales, (2015) <sup>[43]</sup> useful tools in evaluating the effects of pollutants in fish and other aquatic organisms (Ansari *et al.*, 2011) <sup>[44]</sup>. Elimination of amplified genetic materials from the cell (Fench, 2011) <sup>[45]</sup> non incorporated chromosomal breaks or losses during cell division cycle (Renu and Saxena, 2015) <sup>[46]</sup> have been reported to be the output of micronuclei

The present study revealed sub-lethal concentrations of Diuron in that it caused concentration-dependent increase of MN in the red blood cells of *C. gariepinus* as observed in 1.5ml/l after 96 hours. Piancini *et al.*, (2015) <sup>[47]</sup> reported similar observation in *Rhamdia quelen* exposed to atrazine herbicide. Mahboob *et al.*, 2013<sup>[48]</sup> give account of similar observation in *C. gariepinus* exposed to mercury chloride. The increase in MN frequency might be as a result of oxidative stress caused by the production of reactive oxygen specie (ROS). This oxidative stress is caused by the inability of the antioxidant system to eliminate the ROS (Dar *et al.*, 2015<sup>[49]</sup>). The presence of ROS has been reported by Li *et al.*, (2011) <sup>[50]</sup> that it could lead to the destruction of macromolecules, the implication of this is that long exposure of *C. gariepinus* to Diuron can cell damage leading to production micronuclei

### Recommendation

Diuron is of uttermost importance in the control of weed in agricultural farm land it is usually absorbed by the soil and gets into the water body either by runoff or ground water. It also has acute/chronic effect on fish. This study confirmed that exposure to pesticides can result in significant haematological and biochemical changes of fish. This study revealed not only toxicity on fish but also caused micronuclei induction. The indication that Diuron cause haematological alteration and can be mutagenic. Further research is encourage at molecular level to reveal more of the harmful effect of Diuron.

### Acknowledgements.

The authors is grateful to the Department of Fisheries and Aquaculture Technology for releasing her research facilities for the study. The effort of Mr Ope, Biochemistry Department Federal University of Technology in the area of MN assay is also appreciated

### References

1. Bennet WA, Sosa A, Britinger TL. Oxygen tolerance of fathead minnows previously exposed to copper. *Bulletin of Environmental Contamination and Toxicology*. 1995;

- 55(4):517-524.
2. Rahman MZ, Hossain Z, Mellah MFA, Ahmed GU. Effects of diazinon 60EC on *Anabus testudineus*, *Channa punctatus* and *Barbades gomnotus* Naga. The International Center for Living Aquatic Resources Management Quarterly. 2002; 25:8-11.
  3. Giacomazzi S, Cochet N. Environmental impact of diuron transformation: A review. *Chemosphere journal*. 2004; 56(11):1021-1032.
  4. Nussey G, Van Vuren JHJ, Du Preez HH. Effect of copper on the differential white blood cell counts of the Mozambique tilapia (*Oreochromis mossambicus*). *Comparative Biochemistry and Physiology Part C: Pharmacology, Toxicology and Endocrinology*. 1995; 111(3):381-388.
  5. Adhikari S, Sarkar B, Chatterjee A, Mahapatra CT, Ayyappan S. Effect of carboforan on certain haematological parameters and prediction of their recovery in a fresh water teleost, *Labeo rohita* (Hamilton). *Ecotoxicology and Environmental safety*. 2004; 58:220-226.
  6. Maheswaran R, Devapaul A, Muralidharan S, Velmurugan B, Ignacimuthu S. Haematological studies of freshwater fish, *Clarias batrachus* (L.) exposed to mercuric chloride. *International Journal of Integrative Biology*. 2008; 2(1):49-54.
  7. Parveen N, Shadab GGHA. Cytogenetic evaluation of cadmium chloride on *Channa punctatus*. *Journal of Environmental Biology*. 2012; 33:663-666.
  8. Tolga CN, Garanko N, Arkhipchuk VV. Induction of micronuclei and binuclei in blood, gill and liver cells of fishes subchronically exposed to cadmium chloride and copper sulphate. *Food Chemical Toxicology*. 2005; 43:569-574.
  9. Güner U, Muranlı FD. Micronucleus test, nuclear abnormalities and accumulation of Cu and Cd on *Gambusia affinis* (Baird and Girard, 1853). *Turkish Journal of Fisheries and Aquatic Science*. 2011; 11615-622.
  10. Bombail V, Gordon E, Batty J. Application of the comet and micronucleus assays to butterflyfish (*Pholis gunnellus*) erythrocytes from the Firth of Forth, Scotland. *Chemosphere*. 2001; 44:383-392.
  11. Grisolia CK, Cordeiro CMT. Variability in micronucleus induction with different mutagens applied to several species of fish. *Genet Molecular Biology*, 2000; 23:235-239.
  12. Parveen N, Shadab GGHA. Evaluation of micronuclei and haematological profiles as genotoxic assays in *Channa punctatus* exposed to Malathion. *International Journal of Science and Nature*. 2011; 2:625-631.
  13. Mohamed MM, El-Fiky SA, Soheir YM, Abeer AI. Cytogenetic studies on the effect of copper sulfate and lead acetate pollution on *Oreochromis niloticus* fish. *Journal of Cell Biology*. 2008; 3:51-60.
  14. Shah SL, Altinda A. Alterations in the immunological parameters of tench (*Tinca tinca* L.) after acute and chronic exposure to lethal and sublethal treatments with mercury, cadmium and lead. *Turkish Journal of Veterinary and Animal Science*. 2005; 29:1163-1168.
  15. Weatherburn MW. Phenol-hypochlorite reaction for determination of ammonia. *Analytical Chemistry*. 1967; 39:971-974.
  16. Perone RD, Madias NE, Levey AS. Serum creatinine as an index of renal function: New insight into old concepts. *Clinical Chemistry*. 1992; 38:1933-1953.
  17. Allain CC, Poon LS, Chan CS, Richmond W, Fu PC. Enzymatic Determination of Total Serum Cholesterol. *Clinical chemistry*, 1974; 20(4):470-475.
  18. Tietz NW. *Fundamental of clinical chemistry*. Philadelphia: Saunders 1982; 1:263.
  19. Ayllon F, Garcia-Vazquez E. Induction of micronuclei and other nuclear abnormalities in European minnow *Phoxinus phoxinus* and mollie *Poecilia latipinna*: an assessment of the fish micronucleus test. *Mutation Research*. 2000; 467:177-186.
  20. Nwani CD, Nagpure NS, Kumar R, Kushwaha B, Kumar P, Lakra WS. Mutagenic and genotoxic assessment of atrazine-based herbicide to freshwater fish *Channa punctatus* (Bloch) micronucleus test and single cell gel electrophoresis. *Environmental Toxicology and Pharmacology*. 2011; 31:314-322.
  21. Lin CC, Liu CI. Test for ammonia toxicity of cultured hybrid tilapia. In *The Second Asian Fisheries Forum*. Asian Fisheries Society, Manila, Philippines. 1990, 457-459.
  22. Omoniyi I, Agbon A, Sodunk SA. Effects of lethal and sub-lethal concentrations of tobacco (*Nicotiana tobaccum*), leaf dust extractum on weight and haematological changes in *Clarias gariepinus* (Buchell 1822). *Journal of Applied and Environmental Management*. 2002; 6:37-41.
  23. Kulakkattolickal AT, Kramer DI. The Role of Air Breathings in the Resistance of Bimodally Respiring Fish to Water. *J. Fish Biol*. 1997; 32:119-127.
  24. Yunis AJ, Claudiano GS, Marcusso PF, Ikefuti C, Ortega GG, Eto SF, *et al*. Acute toxicity and determination of the active constituents of aqueous extract of *Uncaria tomentosa* Bark in *Hyphessobrycon eques*. *Journal of toxicology*. 2014; 412437:5.
  25. Ukagwu JI, Onuoha GUC, Chude LA. Haematological changes in juvenile catfish (*Clarias gariepinus*) exposed to pulp and paper mill effluent under field condition in Imo river overrrinta, Abia state. *Nigerian journal of Agriculture, food and environment*. 2012; 8(1):86-93.
  26. Barkhordar M, Valizadeh R, Aghili M, Taherimirghaed A, Ghorbani R, Hedayati A. Acute toxicity test of cypermethrine on common carp (*Cyprinus carpio*). *International Journal for Agric Science Research*. 2013; 2(7):234-237.
  27. Fryer JD. *Weed control handbook* Edited by Make Peace, 1977; 1:384-389.
  28. Oropesa AL, Garcia – Cambero JP, Gomez L, Roncero V, Soler V. Effect of long term exposure to simazine on histopathology, hematological, biochemical parameters in *Cyprinus carpio*. *Environmental Toxicology*. 2009; 24:187-199.
  29. Svoboda M, Luskova V, Drastichova J, Žlabek V. The effect of diazinon on haematological indices of common carp (*Cyprinus carpio* L.). *Acta Veterinaria Brno*. 2001; 70(4):457-465.
  30. Olufayo MO. Haematological characteristics of *Clarias gariepinus* (Burchell 1822) juveniles exposed to *Derri elliptica* root powder. *African Journal of Food, Agriculture, Nutrition and Development*. 2009; 9(3):919-933.

31. Ayotunde EO, Fagbenro OA, Adebayo OT, Amoo AI. Toxicity of Aqueous Extracts of Drumstick, *Moringa oleifera*, Seeds to Nile tilapia, *Oreochromis niloticus*, Fingerlings and Adults. In *Proceedings of the 6th International Symposium on Tilapia in Aquaculture*, 2004, 200-208.
32. Rottman RW, Francis-Floyd R, Durborow R. The role of stress in fish disease Southern Regional. *Aquaculture Centre (SRAC), Pub. Publication*, 1992, 474.
33. Larsson Å, Haux C, Sjöbeck ML. Fish physiology and metal pollution: results and experiences from laboratory and field studies. 1985; 9(3):250-281.
34. Adekunle IM, Arowolo TA, Omoniyi IT, Olubambi OT. Risk assessment in Nile tilapia (*Oreochromis niloticus*) and African mud catfish (*Clarias gariepinus*) exposed to cassava effluent. *Chemistry and Ecology*. 2007; 23(5):383-392.
35. Ruparelia SG, Verma Y, Saiyed SR, Rawal UM. Effect of cadmium on blood of tilapia, *Oreochromis mossambicus* (Peters), during prolonged exposure. *Bulletin of environmental contamination and toxicology*. 1990; 45(2):305-312.
36. Murray RK, Granne DK, Mayers PA, Rodwell PW. Harper's Biochemistry 23rd ed., Appleton and Lange Publishers, Norwalk, Connecticut/Los Altos, California, 1990, 221.
37. Lall SB, Das N, Rama R, Peshin SS, Khatter S, Gulati K, Seth SD. Cadmium induced nephrotoxicity in rats. *Indian Journal of Experimental Biology*. 1997; 35:151-154.
38. John PJ. Alteration of certain blood parameters of freshwater teleost *Mystus vittatus* after chronic exposure to metasytox and sevin. *Fish Physiology and Biochemistry*. 2007; 33(1):15-20.
39. Banaee M. Physiological dysfunction in fish after insecticides exposure: Insecticides often undesired but still so Important, Edited by Stanislav Trdan, Published by InTech. 2013; 4:103-142.
40. Ibrahim NA, El-Gamal BA. Effect of diazinon, an organophosphate insecticide on plasma lipid constituents in experimental animals. *Journal of Biochemistry and Molecular Biology*. 2003; 36(5):499-504.
41. Lasram MM, Annabi AB, Elj NE, Selmi S, Kamoun A, El-Fazaa S, *et al.*, Metabolic disorders of acute exposure to malathion in adult Wistar rats. *Journal of Hazardous Materials*. 2009; 163(2-3):1052-1055
42. Acker CI, Nogueira CW. Chlorpyrifos acute exposure induces hyperglycemia and hyperlipidemia in rats. *Chemosphere*. 2012; 89(5):602-608.
43. Mazzeo DEC, Marin-Morales MA. Genotoxicity evaluation of environmental pollutants using analysis of nuclear alterations. *Environmental Science and Pollution Research*. 2015; 22:9796-9806.
44. Ansari RA, Rahman S, Kaur M, Anjum S, Raisuddin SN. In vivo cytogenetic and oxidative stress-inducing effects of cypermethrin in freshwater fish, *Channa punctatus* Bloch. *Ecotoxicology and Environmental Safety*. 2011; 74:150-159.
45. Fench M. The in vitro micronucleus technique. *Mutation Research*. 2011; 455:81-95.
46. Renu C, Saxena K. Genotoxic evaluation of fenvalerate in *Channa punctatus* by micronucleus test. *Indian Journal of Science and Research Technology*. 2015; 3:30-33.
47. Piancini LDS, Santos GS, Tincani FH, Cestari MM. Piscine micronucleus test and the comet assay reveal genotoxic effects of Atrazine herbicide in the neotropical fish *Rhandia quelen*. *Ecotoxicol Environ Contam*. 2015; 10:55-60.
48. Mahboob S, Al-Balwai HFA, Al-Misned F, Ahmad Z. Investigation on the genotoxicity of mercuric chloride to freshwater *Clarias gariepinus*. *Pak Vet J*. 2013; 34:100-103.
49. Dar SA, Yousuf AR, Balkhi M. Assessment of endosulfan induced genotoxicity and mutagenicity manifested by oxidative stress pathways in freshwater cyprinid fish crucian carp (*Carassius carassius* L.). *Chemosphere*. 2015; 120:273-283.
50. Li ZH, Velisek J, Zlabek V, Grabic R, Machova J, Kolarova. JLiP, Randak T. Chronic toxicity of verapamil on juvenile rainbow trout (*Oncorhynchus mykiss*): effects on morphological indices, hematological parameters and antioxidant responses. *J Hazard Mat*. 2011; 185:870-880.