



International Journal of Fisheries and Aquatic Studies

E-ISSN: 2347-5129

P-ISSN: 2394-0506

(ICV-Poland) Impact Value: 5.62

(GIF) Impact Factor: 0.549

IJFAS 2017; 5(4): 138-146

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www.fisheriesjournal.com

Received: 14-05-2017

Accepted: 15-06-2017

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Impact of Methoprene on Spermatogenesis of the Freshwater Crab *Travancoriana schirnerae*

Latha Nadkandi Padmanabhan, Sudha Devi Arath Raghavan and Aswani Ayanath

Abstract

This study investigated the impact of methoprene, a juvenile hormone analog, on spermatogenesis of the freshwater crab *Travancoriana schirnerae*, abundant in the wetlands of Wayanad. Administration of sublethal dose of methoprene demonstrated reduced gonadosomatic indices and considerable degree of changes in the morphology and histoanatomy of the testis leading to partial or total suppression of spermatogenesis during active phase. Histopathological changes include shrinkage and vacuolization of acini, disintegration of basal lamina and germinal epithelium, deformation, disarray and lower chromatin content of primary gonidia and a reduction in number of germ cells. Decline in mature sperm concentration, abnormal morphology and irregular arrangement of sperms was also observed in experimental crabs, in contrast to the fully laden acini with mature spermatozoa arranged in circles of control crabs.

Keywords: Freshwater crab, histology, juvenile hormone analog, methoprene, *Travancoriana schirnerae*

1. Introduction

The potential of juvenile hormone analogs (JHAs) as pest control agents has gained momentum in the last few years. Unlike the environmentally burdensome chemical insecticides which cause immediate death of target insects, JHAs interfere with growth and development, making it impossible for the immature insects to become adults. In adult insects, JHAs hinder the reproductive processes resulting in sterile individuals^[1, 2]. One of the extensively used JHAs is methoprene which has desirable features such as short environmental half-life, low toxicity and minimal impact on non-target organisms (NTOs)^[3, 4]. It is used in the management of both agricultural and non-agricultural pests especially mosquitoes^[5, 6].

Methoprene enters the freshwater environments either by direct application for controlling aquatic pests or indirectly through land drainage or erosion from adjacent methoprene treated agricultural lands^[1, 6]. The biological activity of aqueous solutions of methoprene in the field is affected by sunlight, temperature and microbial action. Repeated application of methoprene in shallow and poorly flushed waters is considered to be hazardous to aquatic biota and the environment. Studies by Kikuchi et al.^[7] and Hershey et al.^[8] have demonstrated barely minor adverse effects of methoprene on aquatic biota while others have shown that methoprene can be quite toxic to non-target aquatic organisms, including insects^[9] and crustaceans^[10].

In non-target crustaceans, studies have indicated that application of JHAs affects growth^[11], survival^[12], metamorphosis^[13], limb regeneration and moulting^[14, 15], energy metabolism and development^[16]. Much concern has been raised in the past regarding the potential for methoprene use in larvicidal activities which may cause huge impact on crustaceans such as shrimps, crabs and lobsters. Celestial and McKenney^[17] showed a significant reduction in the survival of larval stages of the mud crab *Rhithropanopeus harrisi* exposed to methoprene. Horst and Walker^[18] found that methoprene inhibited morphogenesis and shell formation in the postmoult blue crab *Callinectes sapidus*. In the same species, exposure to methoprene at environmental concentrations 2-10 μ M resulted in morbidity, mortality, reduction in the number of successful hatchlings and lethargic behaviour in surviving zoeae^[19].

Methoprene exposure caused a significant reduction in larval metamorphosis in the estuarine grass shrimp *Palaemonetes pugio*^[20] and lethality in the juvenile mysid shrimp *Mysidopsis bahia*^[21]. Methoprene significantly affected energy and testosterone metabolism^[22],

moulting [23] and survival and longevity [24] in non-target crustaceans. Gradoni [25] found that juvenile amphipods are more sensitive to methoprene than adults.

Studies regarding the impact of methoprene on reproductive aspects of non-target macrocrustaceans are limited [11]; majority of the documented works were directed to acute toxicity [26] and most toxicological studies on crustaceans have not addressed cellular effects [27]. Payen and Costlow [28] observed disturbed spermatogenesis in methoprene (Altosid) treated *R. harrisii*. So far there has been no comprehensive study on the impact of methoprene on reproduction in freshwater crabs. Hence, in the present study, an attempt has been made to investigate the histopathological changes induced in the testis of the freshwater crab *Travancoriana schirnerae* on administration of sublethal dose of methoprene. Histological changes not only give an early indication of pollution hazard but also provide useful data on nature and degree of damage to cells and tissues [29].

2. Materials and methods

2.1 Test chemical

Methoprene is a long chain hydrocarbon ester, characterized as an amber or pale yellow liquid with a faint fruity odour [30]. The commercial product, Diacon II® containing 33.6% (S)-methoprene/L as active ingredient (Wellmark International, IL, USA) was used as the source of methoprene in this study. The test solutions were prepared in distilled water just before administration.

2.2 Determination of LD₅₀ for 48 hours

Adult male intermoult crabs (carapace width 4.0-5.0 cm, body weight 22.39-47.56 g; 33.52±7.56 g) were collected from the paddy fields of Ondayangadi, about 5 km northeast of Mananthavady (11.82° N and 76.02° E, altitude 767 m) in Wayanad district of Kerala, India. Fifty crabs were distributed into five groups of ten each. Group I individuals received injections of 50 ppm (S)-methoprene (in 100 µl distilled water) into the body cavity through the arthroal membrane of the third walking leg. Similarly, individuals of Group II, III, IV and V received injections of 100, 150, 200 and 250 ppm (S)-methoprene/100 µl distilled water, respectively. The crabs were kept separately in plastic tubs and observed for 48 h. The mortality in each group was recorded. The LD₅₀ (155 ppm for 48 h) was determined using Probit analysis [31] (Randhawa, 2009). One tenth of the LD₅₀ (15 ppm) was used as the test dose in the present study.

2.3 Experimental design

Adult male intermoult crabs were collected for a period of one

year (March 2014-February 2015) from the same paddy fields mentioned above and acclimatized to laboratory conditions for three days. They were fed with boiled egg and pulses during the acclimation and experimental period.

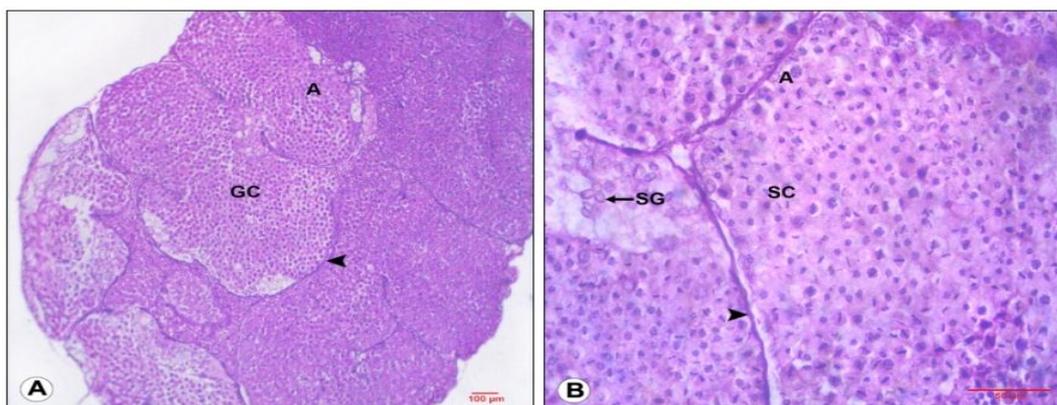
Every month, ten untreated crabs were maintained as controls and another ten as experimentals. To each experimental crab, 100 µl 15 ppm (S)-methoprene was carefully injected into the body cavity through the arthroal membrane of the third walking leg on days 1, 8, 15 and 22. A week after administration of the fourth dose (on day 30), the animals were sacrificed. Testes from experimental and control crabs were carefully dissected out, weighed and fixed in Bouin's solution for histological analyses. The gonadosomatic index (GSI) was calculated using the formula: wet weight of gonad (g)/wet weight of whole body (g) × 100. The tissues were dehydrated in graded series of ethanol and embedded in paraffin wax. Sections 5-6 µm thickness were stained with haematoxylin-eosin and examined under a Leica DM 500 Research Microscope. Measurements were recorded using an image analysis system of Biowizard software. Photomicrographs were taken with a DG 330/210 camera attached to the microscope. Student's t-test was applied in analyzing the data.

3. Results

3.1 Histopathology of testis of methoprene treated crabs during active phase (May-June)

Histological observations of normal, untreated crabs during active phase revealed normal testicular architecture with regular course of spermatogenesis. Acinar walls appeared well stretched as the acini were fully packed with germ cells (Figure 1A, B). The most prominent feature of this phase was that majority of the acini were packed with mature spermatozoa (Figure 1C).

The testes of treated crabs showed abnormal testicular architecture. There occurred a significant reduction in GSI value and acinar diameter when compared to control crabs (Tables 1, 2). Acinar boundaries appeared indistinct in some areas of the testis and degenerative changes were perceptible in the germ cells. Methoprene injection affected the regular, round or oval shape of gonial cells (Table 3). Their nuclei were seen enlarged and vacuolated and chromatin margined along the nuclear membrane was frequently noticed (Figure 1D). In some acini, arrest of division of gonial cells was observed. Primary and secondary spermatocytes were found reduced in size and exhibited nuclear pycnosis preceding degeneration (Table 3). Damaged, elongate spermatids could be seen in a small number of acini.



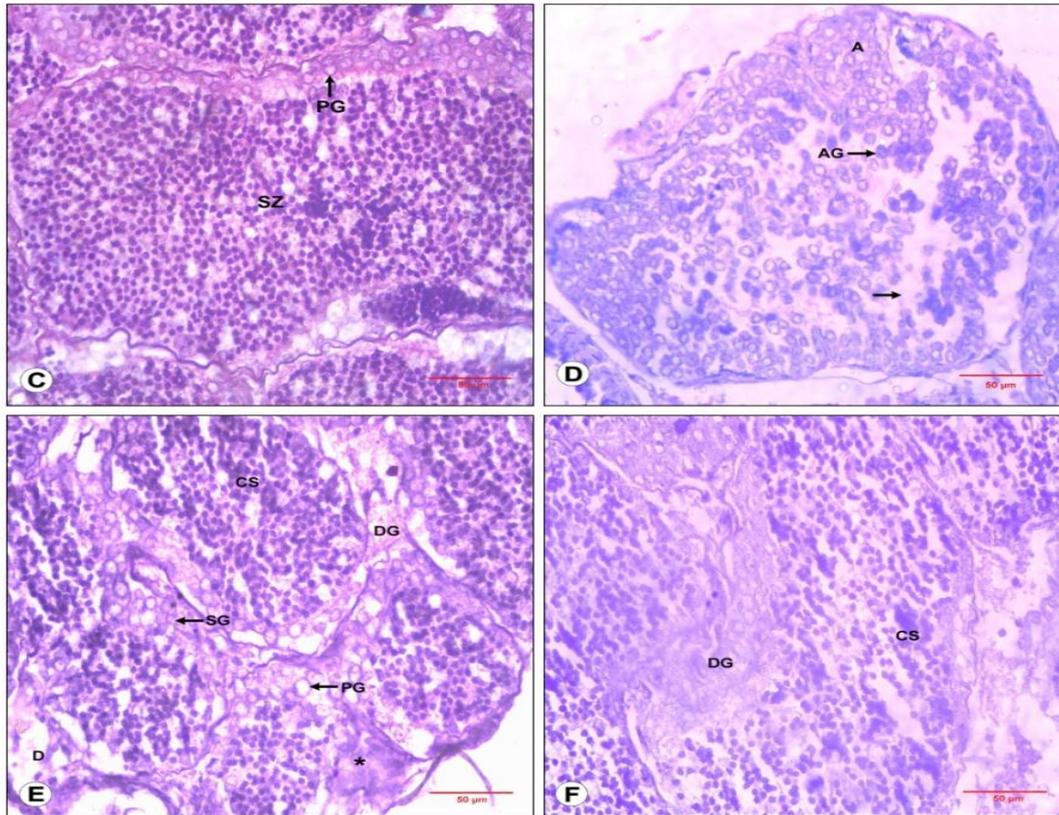


Fig 1: Histological features of the testis of control (A-C) and methoprene treated (D-F) *T. schimerae* during active phase (May-June) of spermatogenesis.

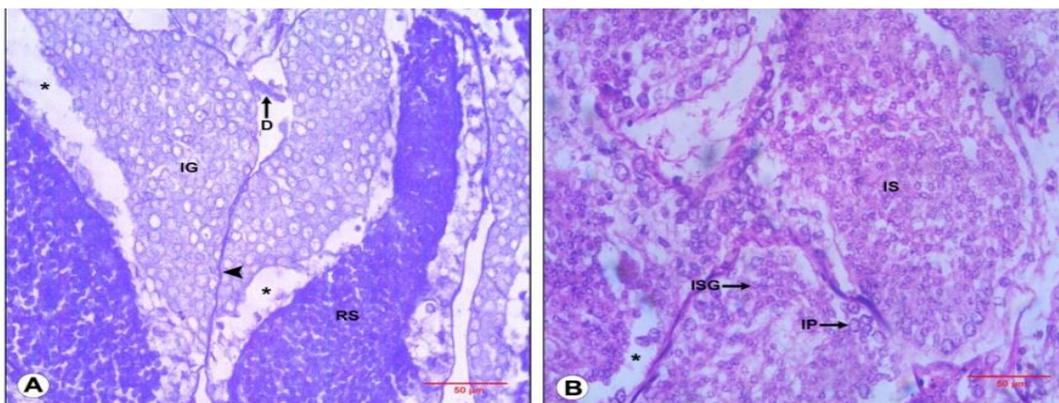
(A) Acini fully packed with active germ cells, (B) Acini with dividing spermatocytes, (C) Mature spermatozoa arranged in circles in acini, (D) Acinus with abnormal gonias and cell free spaces, (E) Acini with clumped sperms and gonias showing signs of degeneration and dissolution, (F) Degenerative changes in the gonial cells and spermatozoa.

A: Acinus; GC: Germ cells; SG: Secondary gonias; SC: Spermatocytes; PG: Primary gonias; SZ: Spermatozoa; AG: Abnormal gonias; CS: Clumped sperms; DG: Dissolved gonias; D: Debris; Arrow head indicates acinar wall; Arrow indicates cell free spaces; Asterisk indicates sperm dissolution. Methoprene treatment appeared to have hampered the development of sperms as indicated by their reduced size (Table 3). Acini packed with sperms arranged in circles (not pycnotic) were still perceptible in some areas of the testis. Many acini carried clumped sperms arranged in the centre and gonias towards the periphery showing signs of degeneration and dissolution (Figure 1E, F). Methoprene injection during this phase caused a significant decrease in sperm count when compared to controls.

3.2 Histopathology of testis of methoprene treated crabs during inactive phase (July-February)

The annual spermatogenic activity began to decline after mating in July-August. The testis remained inactive till February as evidenced by a progressive fall in GSI value and acinar diameter (Tables 1, 2) (Figure 2A-C).

Methoprene treatment hastened the disintegration and degeneration of the germ cells during the inactive phase. When compared to the untreated control crabs, GSI values and acinar diameter of treated crabs reduced drastically (Tables 1, 2). Light microscopy revealed disorganization of acinar elements, extensive inter and intra acinar spaces and cell debris. In some areas of the testis, acinar boundaries were indistinct or broken and the cell types could not be distinguished. The acinar contents appeared inconsistent, clumped, inconspicuous and detached from the wall (Figure 2D). Furthermore, atrophied nature of acini and changes in the basal lamina including irregularity and indentations were observed. Acini showing degenerative changes and appearance of vacuolated germ cells increased.



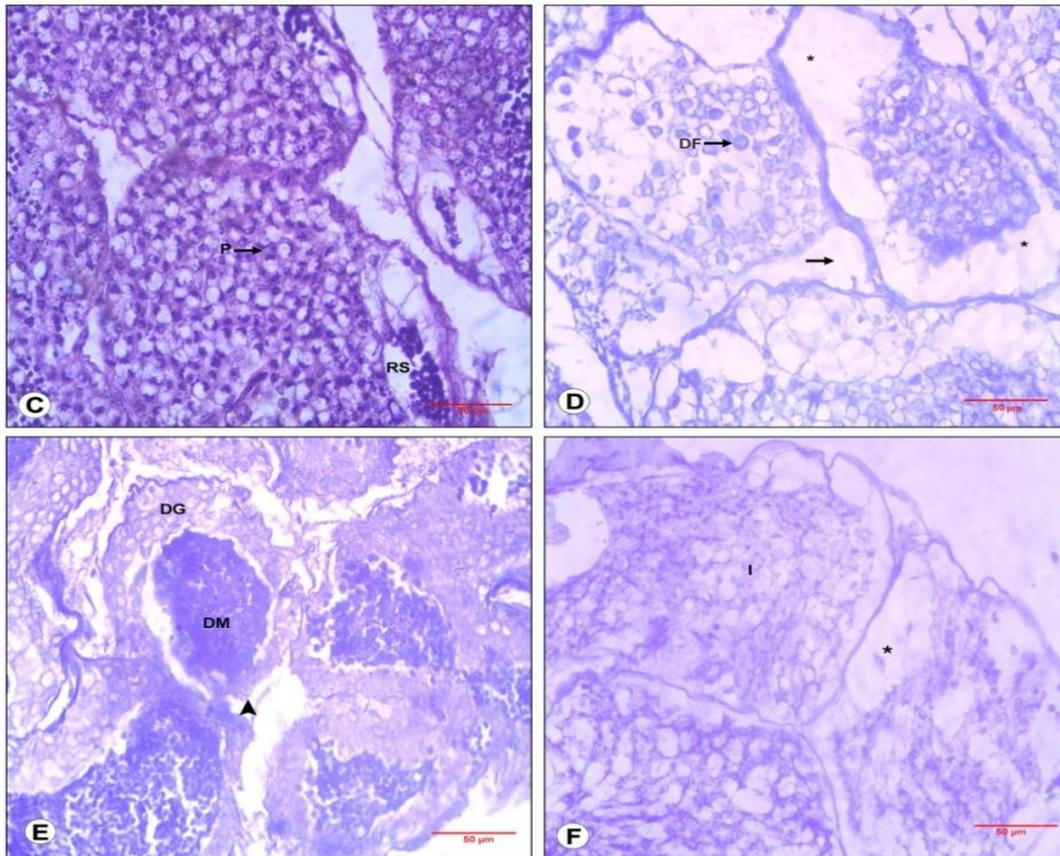


Fig 2: Histology of testis of control (A-C) and treated (D-F) crabs during inactive phase (July-February) of spermatogenesis.

(A) Highly pycnotic residual sperms and inactive gonial cells, (B) Cell free spaces, inactive secondary gonial cells and primary spermatocytes, (C) Acini with pycnotic primary gonial cells and highly condensed residual sperm masses towards the periphery, (D) Detached acinar contents and deformed gonial cells, (E) Degeneration and dissolution of gonial cells and spermatozoa, (F) Dissolved appearance of germ cells in acini.

D: Debris; IG: Inactive gonial cells; RS: Residual sperms; IS: Inactive secondary gonial cells; IP: Inactive primary gonial cells; ISG: Inactive secondary gonial cells; IP: Inactive primary gonial cells; P: Pycnotic gonial cells; DF: Deformed gonial cells; DG: Dissolving gonial cells; DM: Dissolved sperm mass; I: Indistinct germ cells; Arrow head indicates acinar wall; Arrow indicates inter acinar space; Asterix indicates intra acinar space.

Table 1: Gonadosomatic index of control and methoprene treated crabs during different phases of spermatogenesis.

Phase	Months	Control	Experimental
Active	May-Jun	0.181±0.01	0.129±0.01*
Inactive	July-Aug	0.159±0.02	0.117±0.02*
	Sep-Oct	0.138±0.01	0.108±0.01*
	Nov-Dec	0.101±0.02	0.085±0.001*
	Jan-Feb	0.077±0.01	0.069±0.01#
Revival	Mar-April	0.120±0.03	0.103±0.02#

The values are represented as Mean±SD. **P*<0.05, #*P*>0.05.

Histological examination revealed distorted and loosely arranged germ cells giving a dissolved appearance to the germ cells in many acini. Germ cell count was significantly decreased when compared to the control group. Arrested division stages in primary gonial cells, deformation with enlarged and vacuolated nuclei were frequently noticed (Figure 2D). Acini with loosely packed and degenerated gonial cells were abundant. Arrest of division stages and chromatin condensation and necrosis of spermatocytes was noticed in many acini. Spermatids reduced in size, showing nuclear pycnosis and signs of degeneration were frequently encountered in the lumen of some acini. Acini with clumped sperms arranged in the middle or periphery were noticed. A few acini carried degenerating sperms and deformed gonial cells in the periphery showing signs of degeneration and dissolution (Figure 2E). Shrunken acini enclosing clumped and dissolved

sperm masses and reduced intra luminal sperm concentration was perceptible. Pycnotic residual sperm masses undergoing degeneration and dissolution were observed. In some acini, germ cells appeared dissolved and could not be identified (Figure 2F).

Table 2: Changes in the acinar diameter of control and methoprene treated crabs during different phases of spermatogenesis.

Phase	Months	Control (µm)	Experimental (µm)
Active	May-Jun	364.81±12.21	295.10±12.92*
Inactive	July-Aug	241.37±16.38	203.68±3.59*
	Sep-Oct	218.28±2.62	202.40±1.93*
	Nov-Dec	191.68±4.70	172.09±4.33*
	Jan-Feb	173.63±3.06	160.70±6.00*
Revival	Mar-April	194.74±8.54	173.71±12.82*

The values are represented as Mean±S.D. **P*<0.05

3.3 Histopathology of testis of treated crabs during revival phase (March-April)

The annual spermatogenic activity was found revived by March-April, perceptible by the proliferation of gonial cells from the germinal epithelium and pronounced increase in the GSI value and acinar diameter (Tables 1, 2).

Histological observations revealed that methoprene treatment during the revival phase affected the activities of germ cells leading to partial arrest of the testicular processes when compared to the untreated control crabs (Figure 3A-C). The GSI value and the average size of acini decreased significantly (Tables 1, 2).

Methoprene affected the proliferation of gonial cells as their distribution was very much reduced in treated crabs. The regular, round to oval shape of gonial cells was lost and their uniformity was disturbed. Their nuclei appeared shrunken, distorted and chromatin contents were clumped (Figure 3D). Shrunken primary and secondary spermatocytes were noticed. In some acini, dissolved sperm masses were noticed towards the interior while arrested gonial cells were seen towards the periphery (Figure 3E). A few acini carried pycnotic spermatids. Phagocytes were often encountered among the degenerating germ cells. Prominent cell free zones were noticed within and outside the acini (Figure 3F).

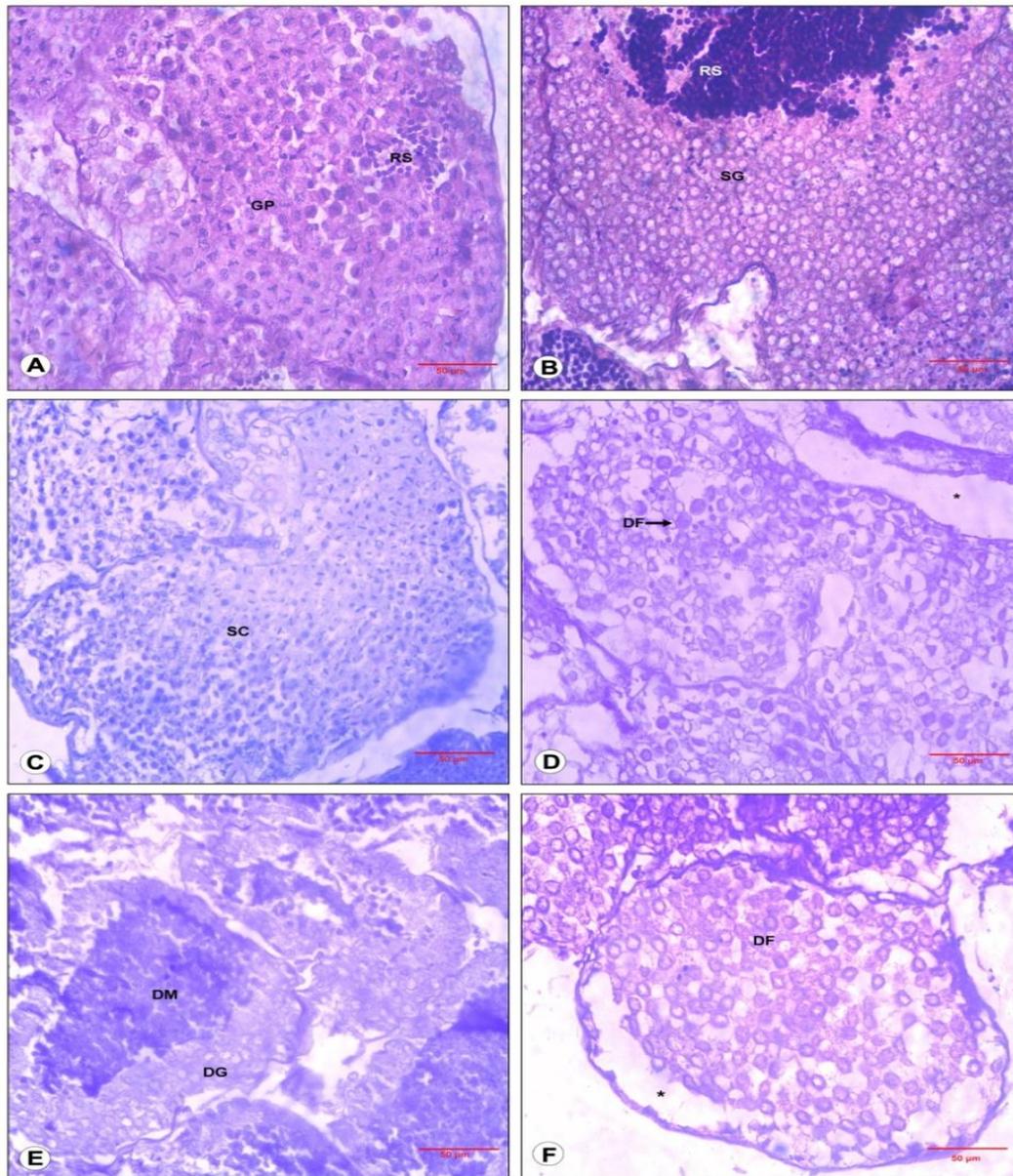


Fig 3: Histology of testis of control (A-C) and methoprene treated (D-F) crabs during revival phase (March-April) of spermatogenesis.

(A) Gonial proliferation, (B) Acini fully packed with secondary gonial cells, (C) Acini with dividing spermatocytes, (D) Deformed gonial cells with shrunken nuclei and clumped chromatin content, (E) Acini with dissolved sperm masses and degenerating gonial cells, (F) Acinus showing deformed gonial cells and cell free spaces.

RS: Residual sperms; GP: Gonial proliferation; SG: Secondary gonial cells; SC: Dividing spermatocytes; DF: Deformed gonial cells; DM: Dissolved sperm masses; DG: Degenerating gonial cells; Asterisk indicates cell free space.

Table 3: Impact of methoprene on the size of germ cells of *T. schirmerae* during active phase of spermatogenesis.

Germ cells	Control (µm)	Experimental (µm)
Primary gonial cells	9.1-12.0	8.2-11.5
Secondary gonial cells	6.1-9.0	4.5-8.2
Primary spermatocytes	5.1-6.0	3.8-4.5
Secondary spermatocytes	3.1-4.0	2.8-4.0
Spermatids	2.0-3.5	2.0-3.0
Spermatozoa	4.0-5.0	3.3-3.8

4. Discussion

The present investigation on acute exposure of sublethal dose of methoprene during different phases of spermatogenesis demonstrated reduced GSI values and considerable degree of changes in the morphology and histoanatomy of the testis leading to partial or total suppression of spermatogenesis.

A reduction in the GSI of methoprene treated *T. schirnerae* was perceptible in all the phases of spermatogenesis. Similar observations on decrease in GSI values were reported in other non-target crustaceans exposed to JHAs, pesticides, heavy metals and pollutants. Significantly low GSI values were recorded in male *Somanniathelphusa pax* collected from arsenic contaminated sites of Hanoi [32, 33]. Wang et al. [34] observed reduced testis weight in the freshwater crab *Sinopotamon henanense* exposed to cadmium. Tributyltin (TBT) exposure exhibited decreased GSI values in the estuarine mud crab *Scylla serrata* and the giant freshwater prawn *Macrobrachium rosenbergii* [35, 36]. On the contrary, oral administration of low and high doses of arsenic did not induce any change in the GSI of male freshwater crab *Geothelphusa dehaani* [33]. Methoprene induced reduction in GSI observed in the present study possibly suggest a reduction in testicular weight caused by degeneration of spermatogenic cells which is an indication of testicular atrophy. Gonadosomatic index reaches its maximum at the peak period of activity and has been considered as a reliable estimate for gonadal maturity of a species [37].

Our results clearly demonstrated that methoprene induced histopathological changes in the testis, such as shrinkage and vacuolization of acini, disintegration of basal lamina and germinal epithelium, deformation, disarray and lower chromatin content of gonidia and reduction in the number of germ cells, were clear indications of atrophic conditions. Our findings were in agreement with the results documented in several non-target crustaceans treated with JHAs, pesticides, heavy metals and pollutants [29, 33, 34]. In *R. harrisii*, long exposure to Altosid caused considerable reduction in spermatogenesis: disappearance of gonidia, stoppage of gonial mitosis and spermatocyte meiosis, presence of many spermatocytes and few spermatids [28]. In the freshwater crab *Paratelphusa hydrodromous*, sublethal concentration of cadmium brought about prominent histological variations in the testicular tissue as indicated by reduction in the number of germ cells, increased size of lumen of seminiferous tubules, disorganization of tubules and lower chromatin content of the proliferating cells and germinal epithelial cells [38].

Light microscopic examination of the testes of control *S. olivacea* showed normal testicular architecture while cadmium treatment resulted in vacuolar degeneration of spermatogenic epithelium, atrophy of seminiferous tubules, edema in the interstitial space as well as extensive necrosis associated with impaired spermatogenesis [39]. Treatment with different concentrations of cadmium induced major histopathological defects such as lesioned germ cells, edema in the interstitial space and extensive germinal layer necrosis in the testis of *S. henanense* [34]. Oral administration of high dose of arsenic in *G. dehaani* illustrated only spermatogonia and spermatocytes while spermatogenesis progressed till spermatozoa stage in the untreated control crabs [33].

Treatment with organophosphate and organochlorine pesticides caused swollen and degenerated testicular layer, vacuolization of acinar contents, reduction in spermatogenic material and disappearance of gonial cells leading to total disintegration of testicular processes in *S. serrata* [40].

Histopathological changes observed in the testicular tissue in *M. lamerrii* and *M. kistensis* include breakdown of germinal epithelium, deformed tubules, hindered gonial proliferation, vacuolization, degeneration of germ cells, damaged proliferation zone and reduced spermatogenic mass on exposure to fenitrothion and phenol respectively [41, 42, 43]. On the contrary, Economopoulos and Gordon [44] did not find any modification of spermatocyte differentiation in the nymphs of the large milkweed bug *Oncopeltus fasciatus* treated with JHAs. Landa and Metwally [45] observed normal spermatogenesis in pupae of Coleoptera treated with a JHA. Szollosi [46] did not observe direct effects on spermatogenesis in late larvae of acridids treated with JHAs like synthetic Cecropia juvenile hormone or hydroprene.

The current study described a decline in mature sperm concentration and irregular arrangement of sperms in acini of treated crabs during the active phase in contrast to acini fully laden with mature spermatozoa arranged in circles of control crabs. Concomitant findings have been reported in other non-target crustaceans exposed to pesticides and pollutants. In *S. henanense*, treatment with different concentrations of cadmium reduced sperm counts adversely affecting male fertility [34]. Different concentrations of TBT yielded abundance of immature sperm cells and a decline in the number of mature spermatozoa in *M. rosenbergii* [47]. Yang et al. [48] found that sperm counts were 20% lower in the amphipod *Echinogammarus marinus* in sites polluted with poly-chlorinated biphenyls, metals and hydrocarbons.

The present investigation also described abnormal sperm morphology as a consequence of methoprene administration during the active phase. Wang et al. [49] found ultrastructural changes in sperm cells with damaged acrosome, mitochondria, endoplasmic reticulum and cell membranes in cadmium injected *S. yangtsekiense*. Decreased sperm length in *M. rosenbergii* exposed to TBT was documented by Revathi et al. [47]. A decline in the concentration, irregular arrangement and abnormal morphology of mature spermatozoa in acini of treated crabs of the present study can be ascribed to the deformation and degeneration of germ cells induced by methoprene. McLachlan et al. [50] have commented that the reduction in sperm quantity or quality may be caused by alterations in testicular development, intrinsic defects in the ability of germ cells to divide and differentiate or impacts on the hormonal regulatory pathway.

In *T. schirnerae*, sensitivity of the testis to methoprene changed depending on the phase of spermatogenesis, i.e. the testis was more sensitive to methoprene during the active and revival phases and less sensitive during the inactive phase. The reduction in GSI values in addition to histopathological changes was more evident in the active and revival phases when compared to the inactive phase. Yamaguchi et al. [33] reported that the sensitivity of the testicular tissue to arsenic changes depending on the season or the developmental stage of the testis in *S. pax*. On the other hand, Gangotri and Matkar [51] revealed damaged testicular layer and vacuolations in testicular tubules of the freshwater crab *Barytelphusa guerini* subjected to acute and chronic treatment of sugar industry effluents during breeding and non-breeding seasons.

The exact mechanism of methoprene induced histopathological changes in the testis of *T. schirnerae* is unclear. There are two possibilities: one is direct toxicity of methoprene on germ cells which include gonidia, spermatocytes, spermatids and spermatozoa, thus negatively impacting spermatogenesis and male fertility in turn. Another

possibility is that methoprene disrupts the endocrine system, continuously inhibiting spermatogenesis by causing a reduction in the circulating gonad stimulating hormone (GSH) titres which in turn inhibits the secretion of androgenic gland hormone (AGH). Spermatogenesis in crustaceans is regulated by GSH from the brain and thoracic ganglion^[52, 53] and AGH from the androgenic gland (AG). The AGH has been shown to play a role in the regulation of male sex differentiation and spermatogenesis in crustaceans^[54, 55]. In the present study, methoprene injections might have hindered the activity of the AG eventually causing malfunctioning of the testicular cells. According to Payen and Costlow^[28], inhibition of testicular processes by Altosid in *R. harrisii* indicates a chemosterilizing action of methoprene, either working with methyl farnesoate or outcompeting it for the same active site. Several authors reported the adverse effects of insecticides on gonad development and maturation and steroidogenesis in NTOs^[43, 56, 57]. Sastry and Miller^[58] reported methoprene induced destabilization of intracellular lysosomes which result in cellular damage. Mercier et al.^[59] suggested that the histopathological changes in the gonad is due to the probable loss of biochemical reserves reflected by thinning of germinal epithelium as well as poor development of gonad.

Conclusion

Acute exposure of sublethal dose of methoprene during different phases of spermatogenesis of *T. schirmerae*, a non-target crustacean, demonstrated reduced GSI values and considerable degree of changes in the morphology and histoanatomy of the testis leading to partial or total suppression of spermatogenesis. Results of this study provide evidence that the reproductive physiology of non-target organisms is at risk, in spite of methoprene being a comparatively safer option in pest control.

Acknowledgements

This work was supported by Kannur University under the Junior Research Fellowship grant to the first author. The authors are thankful to Dr Paul G Fields, Cereal Research Centre, Manitoba, Canada for the kind gift of Diacon II (methoprene).

References

- Retnakaran A, Granett G, Ennis T. Insect Growth Regulators. In: Comprehensive Insect Physiology, Biochemistry and Pharmacology (eds. by G.A. Kerkut and L.I. Gilbert), Volume 12. Oxford, Pergamon, 1985.
- Oberlander H, Silhacek DL. Insect Growth Regulators. In: Alternatives to Pesticides in Stored-Product IPM (ed. by B.H. Subramanyam and D.W. Hagstrum), Kluwer Academic Publishers, Norwell. 2000, 147-163.
- Ankley GT, Tietge JE, DeFoe D, Jensen GK, Holcombe K, Durhan GW *et al.* Effects of ultraviolet light and methoprene on survival and development of *Rana pipiens*. Environmental Toxicology and Chemistry. 1998; 17:2530-2542.
- Lawler SP, Dritz DA, Jensen T. Effects of sustained-release methoprene and combined formulation of liquid methoprene and *Bacillus thuringiensis israelensis* on insects in salt marshes. Archives of Environmental Contamination and Toxicology. 2000; 39:177-182.
- Hodek I, Ruzicka Z, Sehnal F. Termination of diapause by juvenoids in two species of ladybirds (Coccinellidae). Experientia. 1973; 29:1146-1147.
- Dhadialla TS, Carlson GR, Le DP. New insecticides with ecdysteroidal and juvenile hormone activity. Annual Review of Entomology. 1998; 43:545-569.
- Kikuchi T, Kamel M, Okubo S, Yasuno M. Effects of the insect growth regulator methoprene and organophosphorus insecticides against non-target organisms in urban drains. Japanese Journal of Sanitary Zoology. 1992; 43:65-70.
- Hershey AE, Shannon L, Axler R, Ernst C, Mickelson P. Effects of methoprene and *Bti* (*Bacillus thuringiensis* var. *israelensis*) on non-target insects. Hydrobiologia. 1995; 308:219-227.
- Gelbič I, Papáček M, Pokuta J. The effects of methoprene S on the aquatic bug *Ilyocoris cimicoides* (Heteroptera, Naucoridae). Ecotoxicology. 1994, 1986; 3:89-93.
- Christiansen ME, Costlow JD, Monroe RJ. Effects of the juvenile hormone mimic ZR-515 (Altosid®) on larval development of the mud crab *Rhithropanopeus harrisii* in various salinities and cyclic temperatures. Marine Biology. 1977; 39:269-279.
- Olmstead AW, Le Blanc GL. Low exposure concentration effects of methoprene on endocrine-regulated processes in the crustacean *Daphnia magna*. Toxicological Sciences. 2001; 62:268-273.
- Walker AN, Bush P, Puritz J, Wilson T, Chang ES, Miller Y *et al.* Bioaccumulation and metabolic effects of the endocrine disruptor methoprene in the lobster, *Homarus americanus*. Integrative and Comparative Biology. 2005; 45:118-126.
- Gomez ED, Faulkner DJ, Newman WA, Ireland C. Juvenile hormone mimics: effect on cirriped crustacean metamorphosis. Science. 1973; 179:813-814.
- Stueckle TA, Likens J, Foran CM. Limb regeneration and moulting processes under chronic methoprene exposure in the mud fiddler crab, *Uca pugnax*. Comparative Biochemistry and Physiology. 2008; 147C:366-377.
- Stueckle TA, Shock B, Foran CM. Multiple stressor effects of methoprene, permethrin and salinity on limb regeneration and molting in the mud fiddler crab (*Uca pugnax*). Environmental Toxicology and Chemistry. 2009; 28:2348-2359.
- Tuberty SR, McKenney CL. Ecdysteroid responses of estuarine crustaceans exposed through complete larval development to juvenile hormone agonist insecticides. Integrative and Comparative Biology. 2005; 45:106-117.
- Celestial DM, Mc Kenney CL. The influence of an insect growth regulator on the larval development of the mud crab *Rhithropanopeus harrisii*. Environmental Pollution. 1994; 85:169-173.
- Horst MN, Walker AN. Effects of the pesticide methoprene on morphogenesis and shell formation in the blue crab, *Callinectes sapidus*. Journal of Crustacean Biology. 1999; 19:699-707.
- Glare TR, O'Callaghan M. Environmental and health impacts of insect juvenile hormone analogue, S-methoprene. Report for the New Zealand Ministry of Health, 1999.
- McKenney CL, Mathews E. Influence of an insect growth regulator on the larval development of an estuarine shrimp. Environmental Pollution. 1990; 64:169-178.
- McKenney CL, Celestial DM. Modified survival, growth and reproduction in an estuarine mysid (*Mysidopsis*

- bahia*) exposed to a juvenile hormone analogue through a complete life cycle. *Aquatic Toxicology*. 1996; 35:11-20.
22. Verslycke T, Fockede N, McKenney CL, Roast SD, Jones MB, Mees J *et al.* Mysid crustaceans as potential test organisms for the evaluation of environmental endocrine disruption: a review. *Environmental Toxicology and Chemistry*. 2004; 23:1219-1234
 23. Ghekiere A, Verslycke T, Fockede N, Janssen CR. Non-target effects of the insecticide methoprene on molting in the estuarine crustacean *Neomysis integer* (Crustacea: Mysidacea). *Journal of Experimental Marine Biology and Ecology*. 2006; 332:226-234.
 24. Chu KH, Wong CK, Chiu KC. Effects of the insect growth regulator (S)-methoprene on survival and reproduction of the freshwater cladoceran *Moina macrocopa*. *Environmental Pollution*. 1997; 96:173-178.
 25. Gradoni L, Bettini S, Majori G. Toxicity of Altosid to the crustacean *Gammarus aequicauda*. *Mosquito News*. 1976; 36:294-297.
 26. Meyers TR, Hendricks JD. Histopathology. In: *Fundamentals of Aquatic Toxicology* (eds. by G.M. Rand and S.R. Petrocelli), New York: Hemisphere. 1985; 283-331.
 27. De Fur PL, Crane M, Ingershold C, Tattersfield L. Endocrine disruption in invertebrates: endocrinology, testing and assessment. Pensacola: Society of Environmental Toxicology and Chemistry. 1999.
 28. Payen GG, Costlow JD. Effects of a juvenile hormone mimic on male and female gametogenesis of the mud crab *Rhithropanopeus harrisi* (Gould) (Brachyura: Xanthidae). *The Biological Bulletin*. 1977; 152:199-208.
 29. Shaikh FI, Ustad IR, Ansari NT. Effect of mercuric chloride on the ovary of freshwater crab *Barytelphusa cunicularis*. *The BioScan*. 2010; 5:335-338.
 30. *Farm Chemicals Handbook*. Farm Chemicals Handbook, Volume 83. Meister Publishing Company. 1997.
 31. Randhawa MA. Calculation of LD₅₀ values from the method of Miller and Tainter (1944). *Journal of Ayub Medical College Abbottabad*. 2009; 21:184-185.
 32. Yamaguchi S, Ito A, Higashino T, Miura C, Agusa T, Kubota R *et al.* Influence of endocrine disruptors on reproduction of aquatic animals in Indo-China. *Proceedings of the International Symposium on the development of water resource management system in Mekong watershed*, December, Hanoi, Vietnam, 2004, 32-40.
 33. Yamaguchi S, Celino FT, Ito A, Agusa T, Tanabe S, Bui Cach T *et al.* Effects of arsenic on gonadal development in freshwater crab, *Somanniathelphusa pax* in Vietnam and *Geothelphusa dehaani* in Japan. *Ecotoxicology*. 2008; 17:772-780.
 34. Wang L, Xu T, Lei W, Liu D, Li Y, Xuan R *et al.* Cadmium-induced oxidative stress and apoptotic changes in the testis of freshwater crab, *Sinopotamon henanense*. *PLoS ONE*. 2011; 6:e27853.
 35. Revathi P, Iyapparaj P, Arockia Vasanthi L, Munuswamy N, Krishnan M. Bioaccumulation of tributyltin and its impact on spermatogenesis in mud crab *Scylla serrata* (Forsk.) *Turkish Journal of Biology*. 2013a; 37:296-304.
 36. Revathi P, Iyapparaj P, Arockia Vasanthi L, Munuswamy N, Krishnan M. Ultrastructural changes during spermatogenesis, biochemical and hormonal evidences of testicular toxicity caused by TBT in prawn *Macrobrachium rosenbergii* (De Man). *Environmental Toxicology*. 2013b; 29:1171-1181.
 37. Utkal K, Mohanty S, Dasgupta S, Sahu AK. Effluent on glycogen, protein and free amino acid on the tissue of fish. *Journal of the Inland Fisheries Society of India*. 2010; 42:8-13.
 38. Padmanabhan AM, Mohan K. Sublethal effects of cadmium on testicular and ovarian maturation in freshwater crab *Paratelphusa hydrodromous*. *International Journal of Pharmaceutical Sciences Review and Research*. 2013; 23:43-46.
 39. Rani DS, Kavitha R, Padmaja M. Histological and biochemical changes in reproductive organs of mud crab *Scylla olivacea* (Herbst, 1796) exposed to cadmium nanoparticle. *Journal of Academia and Industrial Research*. 2013; 2:391-396.
 40. Rao KS. Effect of pesticide pollutants on reproduction and neurosecretion in marine edible crab *Scylla serrata*. PhD, Dr Babasaheb Ambedkar Marathwada University, India, 1984.
 41. Mary A. Effect of pesticide on physiology of the freshwater prawn *Macrobrachium lamerrii*. PhD, Dr Babasaheb Ambedkar Marathwada University, India, 1984.
 42. Gangshettiwar VB. Effect of phenol poisoning on the physiology of the prawn, *Macrobrachium lamerrii*. PhD, Dr Babasaheb Ambedkar Marathwada University, India, 1986.
 43. Dode CR, Chourpagar AR, Nagabhushanam R. Histological alteration in testis of the freshwater prawn, *Macrobrachium kistensis* exposed to cuprous oxide. *App Res Dev Inst J*. 2013; 8:9-16.
 44. Economopoulos AP, Gordon HT. Growth and differentiation of testes in the large milkweed bug *Oncopeltus fasciatus* (Dallas). *Journal of Experimental Zoology*. 1971; 177:391-405.
 45. Landa V, Metwally MM. Effects of conventional chemosterilants and juvenile hormone analogues on spermatogenesis in *Trogoderma granarium* (Coleoptera, Dermestidae). *Acta Entomol Bohemoslov*. 1974; 71:145-152.
 46. Szollosi A. Imaginal differentiation of the spermduct in acridids: effects of juvenile hormone. *Acrida*. 1975; 4:205-216.
 47. Revathi P, Iyapparaj P, Arockia Vasanthi L, Munuswamy N, Prasanna VA. Influence of short term exposure of TBT on the male reproductive activity in freshwater prawn *Macrobrachium rosenbergii* (De Man). *Bulletin of Environmental Contamination and Toxicology*. 2014; 93:446-451.
 48. Yang G, Kille P, Ford AT. Infertility in a marine crustacean: have we been ignoring pollution impacts on male invertebrates? *Aquatic Toxicology*. 2008; 8:81-87.
 49. Wang L, Sun HF, Li CY. Effects of cadmium on spermatogenesis in freshwater crab (*Sinopotamon yangtsekiense*). *Acta Zoologica Sinica*. 2002; 48:677-684.
 50. McLachlan JA, Newbold RR, Li S, Negishi M. Are estrogens carcinogenic during development of the testes? *APMIS*. 1998; 106:243-244.
 51. Gangotri MS, Matkar LS. Tissue histopathology of the freshwater crab, *Barytelphusa guerini* (H. Milne Edwards) (Decapoda: Potamidea) under sugar industrial effluent toxicity. *Ecology, Environment and*

- Conservation Papers. 2004; 10:27-36.
52. De Kleijn DPV, Van Herp F. Involvement of the hyperglycemic neurohormone family in the control of reproduction in decapods crustaceans. *Invertebrate Reproduction and Development*. 1998; 33:263-272.
 53. La Font R. The endocrinology of invertebrates. *Ecotoxicology*. 2000; 9:41-57.
 54. Charniaux-Cotton H. Découverte chez un Crustacé Amphipode (*Orchestia gammarella*) glande endocrine responsable de la différenciation de caractères sexuels primaires et secondaires males. *Comptes Rendus de l'Académie des Sciences*. 1954; 239:780-782.
 55. Okumura T, Hara M. Androgenic gland cell structure and spermatogenesis during the moult cycle and correlation to morphotypic differentiation in the giant freshwater prawn, *Macrobrachium rosenbergii*. *Zoological Science*. 2004; 21:621-628.
 56. Dutta HM, Meijer HJM. Sublethal effects of diazinon on the structure of the testis of bluegill, *Lepomis macrochirus*: a microscopic analysis. *Environmental Pollution*. 2003; 125:355-360.
 57. Banaee M, Mirvaghefi AR, Ahmadi K, Ashori R. The effect of diazinon on histopathological changes of testis and ovaries of common carp (*Cyprinus carpio*). *Journal of Marine Biology*. 2009; 1:25-35.
 58. Sastry AN, Miller DD. Application of biochemical and physiological responses to water quality monitoring in biological monitoring of marine pollutants. In: Vernberg FJ, Calabrese A, Thunberg FP, Vernberg WB, editors. London: Academic Press, 1981.
 59. Mercier A, Pelletier E, Hamel J. Metabolism and subtle toxic effects of butyltin compounds in starfish. *Aquatic Toxicology*. 1994; 28:259-273.